

## Release of Nutrients from On-site Wastewater Disposal Systems, Point Pelee National Park, Ontario, Canada\*

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### Abstract

Point Pelee National Park relies on the use of on-site methods to dispose of wastewater generated in the Park. The Park is located on the southern half of a 15 km long cusplate, consisting of two narrow barrier bars and a large interior protected marsh. There are more than 30 active tile beds, excavated directly into the native sands of the western barrier bar. These tile beds receive wastewater from more than 0.5 million day-visitors each year. Past methods of sewage disposal for up to one million day-visitors each year included direct disposal to the barrier bar, the use of holding tanks, and dispersed discharge through tile lines. Because of limited land mass, wastewater disposal often occurred close to the marsh. At two recent tile bed sites, located within 60 m of the marsh, concentrations of NO<sub>3</sub> between 1 and 80 mg/L (as N) are present more than 40 m from the tile beds. In reducing zones at the marsh edge and at the base of the aquifer, bacterial denitrification processes lead to the removal of NO<sub>3</sub> to concentrations < 0.01 mg/L N. Phosphate concentrations approach 3 mg/L P up to 10 m from the tile beds, and 0.1 - 1 mg/L P in a reducing zone at the base of the aquifer up to the marsh edge. Elevated concentrations of NH<sub>3</sub> are present more than 60 m from the tile beds. At a site which last received sewage input two decades ago, NO<sub>3</sub> is absent, but elevated concentrations of PO<sub>4</sub> and NH<sub>3</sub> are present in the groundwater zone close to the marsh edge. Monitoring at the groundwater/marsh interface indicates that nutrient-rich groundwater is seeping into the marsh for most of the year. The eutrophic conditions observed in several marsh ponds adjacent to the barrier bar are attributed to PO<sub>4</sub> release from recent and past on-site wastewater disposal practices.

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### Introduction

The Point Pelee marsh is a major protected coastal wetland located on the north shore of Lake Erie (Figure 1). Several ponds within the marsh experience prolific algal blooms as a result of excess nutrient concentrations (total P > 0.3 mg/L) in the marsh water column (McCrea, 1993). At the end of the growth cycle, the biomass dies, settles to the marsh bottom, and decay processes begin. Mayer et al. (1998) reported elevated concentrations of nutrients in the upper 30 cm of marsh sediments below a hypereutrophic marsh pond. These nutrients, released from anaerobic decay processes, occur as dissolved phosphate and ammonia in

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interstitial pore waters and as accumulated phosphate solids. A portion of these nutrients re-enter the water column and are available for renewed algal growth.

Figure 1. Map of Point Pelee, Ontario, showing locations of western barrier bar, interior marsh, and Blue Heron, Camp Henry and Marsh-Boardwalk sites.

This study was conducted to determine whether input of sewage-derived nutrients is a potential cause of the elevated nutrient concentrations observed in the marsh ponds and sediments. Point Pelee National Park, like many parks in Ontario, relies on the use of septic-tanks and tile bed leach fields to dispose of sewage generated in the Park. Sewage released to the subsurface from tile beds typically leads to the development of plumes of groundwater which contain elevated concentrations of nutrients and other dissolved constituents. These plumes are often many tens of meters in length, and, in some cases, discharge to surface water bodies.

At Point Pelee, there are more than 30 active tile beds, including conventional and raised beds, installed in the western barrier bar of the Park (Figure 1). These tile beds receive wastewater seasonally or year-round from comfort stations, the visitors' centre and Park support buildings. The Park receives approximately 0.5 million visitors each year. In the 1950s and 60s, the number of visitors to the Park was higher, approaching 1 million visitors each year. There were also a number of private dwellings which have been removed through a land acquisition and naturalization program.

The presence and persistence of sewage-derived components in the subsurface was evaluated by conducting hydrogeological studies at active tile beds, a tile bed which had been decommissioned for two years, and a site which last received wastewater more than two decades ago. Measurements of groundwater seepage and nutrient transport directly into the Point Pelee marsh were also made. These measurements assist in the determination of the current rate of release of nutrients to the marsh, and the potential for future release as a result of ongoing discharge of sewage.

## **Study Methods**

Networks of stand-pipe and multilevel bundle piezometers were installed at the Blue Heron and Camp Henry active tile bed sites, and the formerly used Marsh-Boardwalk site (Figures 1 and 2). The Blue Heron tile bed receives about 10,000 L/day of blackwater year-round from a visitor comfort station. It was installed around 1980. The Camp Henry "old" tile bed received blackwater seasonally from an overnight camp for about 17 years. It was sampled while it was active, and two years after it was decommissioned. The Marsh-Boardwalk area had a number of cottages and other buildings, vault toilets, and possibly a tile bed. Buildings at the site were removed and wastewater disposal in the area ceased in the late 1970s. Samples of groundwater were collected from more than 300 locations at the three sites and analysed for concentrations of nutrients ( $\text{NO}_3$ ,  $\text{NH}_3$ ,  $\text{PO}_4$ , and DOC), major ions, and trace metals, and field pH, Eh, alkalinity and temperature. Groundwater flow directions and rates were measured at the groundwater/marsh interface using nests of minipiezometers and seepage meters (Figure 2).

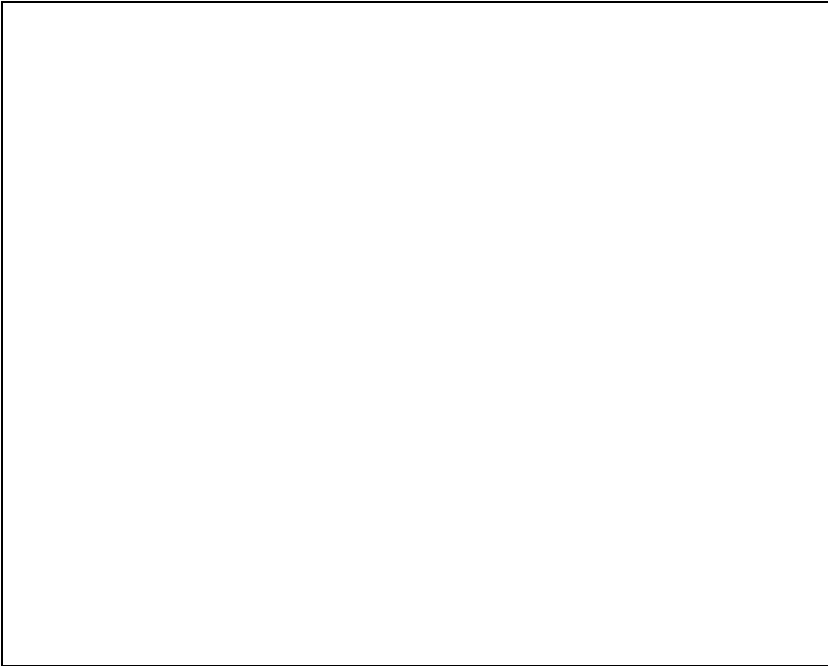


Figure 2. Schematic diagram showing instrumentation techniques used in this study

### **Active Tile Beds**

The composition of effluent entering the active beds is typical of blackwater. It contains elevated concentrations of dissolved organic carbon (DOC), nitrogen, phosphorous, and pathogens (Table 1). When effluent is discharged to tile beds and allowed to infiltrate to the subsurface, a series of reactions occurs. In the unsaturated zone above the water table, DOC and ammonia ( $\text{NH}_3$ ) are oxidized to carbon dioxide and nitrate ( $\text{NO}_3$ ). Organic phosphorous and phosphate ( $\text{PO}_4$ ) are removed through adsorption and precipitation reactions, and pathogens are removed through a variety of physical and chemical processes. If the residence time in the unsaturated zone is sufficient, large declines in concentrations of DOC and ammonia are observed, and some removal of  $\text{PO}_4$  and pathogens occurs. Wastewater entering the saturated zone, however, typically has elevated concentrations of  $\text{NO}_3$  and  $\text{PO}_4$ , and potentially elevated concentrations of  $\text{NH}_3$ , bacteria, viruses and trace contaminants

At the instrumented active sites, elevated concentrations of  $\text{NO}_3$ ,  $\text{PO}_4$  and  $\text{NH}_3$  were observed in the groundwater below and adjacent to the tile beds (Figures 3 and 4) (Ptacek et al., 1994; 1997; Ptacek, 1998). At the Blue Heron tile bed, concentrations of  $\text{NO}_3$  ranged between 1 and 30 mg/L (as N) in a plume that extends more than 40 m from the edge of the bed. At Camp Henry, concentrations of  $\text{NO}_3$  were higher, ranging between 1 and 80 mg/L in a plume that extends more than 40 m from the edge of the "old" bed. These

concentrations are well in excess of the World Health Organization (WHO) drinking water guideline of 10 mg/L  $\text{NO}_3\text{-N}$ . At both sites, in reducing zones at the marsh edge and at the base of the aquifer, bacterial denitrification processes result in the removal of  $\text{NO}_3^-$  to concentrations  $< 0.01$  mg/L N. At both sites, concentrations of  $\text{NH}_3$  are elevated to distances more than 60 m from the tile beds. These elevated concentrations of  $\text{NH}_3$  suggest oxidation of the septic tank effluent was incomplete during its transport through the unsaturated zone. Phosphate concentrations exceed 1 mg/L P up to 10 m from both beds. At the base of the aquifer at both sites, in a reducing zone, concentrations of  $\text{PO}_4\text{-P}$  range between 0.1 - 1 mg/L, up to the marsh edge. These concentrations represent a large decline from approximately 5 - 10 mg/L in the original effluent (Table 1). The large volume of groundwater containing concentrations  $> 0.1$  mg/L P, however, represents sufficient  $\text{PO}_4$  to be of concern if discharged to surface water bodies. Similarly, elevated concentrations of  $\text{NH}_3$  represent a potential input of N into the marsh.

Parameter	Concentration	
	Camp Henry	Blue Heron
$\text{NO}_2+\text{NO}_3$ (mg/L as N)	0.05*	$<0.05$
$\text{NH}_3$ (mg/L as N)	97.9	36.4
P, total (mg/L)	11.8	4.12
DOC (mg/L)	31.8	34.7

\* Concentration equal to analytical detection limit

Table 1: Composition of septic-system effluent collected from Camp Henry and Blue Heron holding tanks, May, 1996.

### Recently-Decommissioned Tile Bed

In 1995, wastewater discharge at the Camp Henry site was switched to a new "raised" tile bed to enhance oxidation of the wastewater. The plume at the "old" tile bed was sampled two years after input had been terminated to evaluate the rate of dissipation of the sewage at the site. This sampling showed large declines in concentrations of  $\text{NO}_3^-$  and  $\text{NH}_3$ , but virtually no change in  $\text{PO}_4$  concentrations (Figure 4). Analyses of  $\text{PO}_4$  conducted on core material collected from the site while sewage disposal was active indicated substantial accumulations of  $\text{PO}_4$  were present on the aquifer solids. The elevated concentrations of  $\text{PO}_4$  in the groundwater are consistent with release of  $\text{PO}_4$  into the flowing groundwater from these earlier accumulations on the aquifer solids.

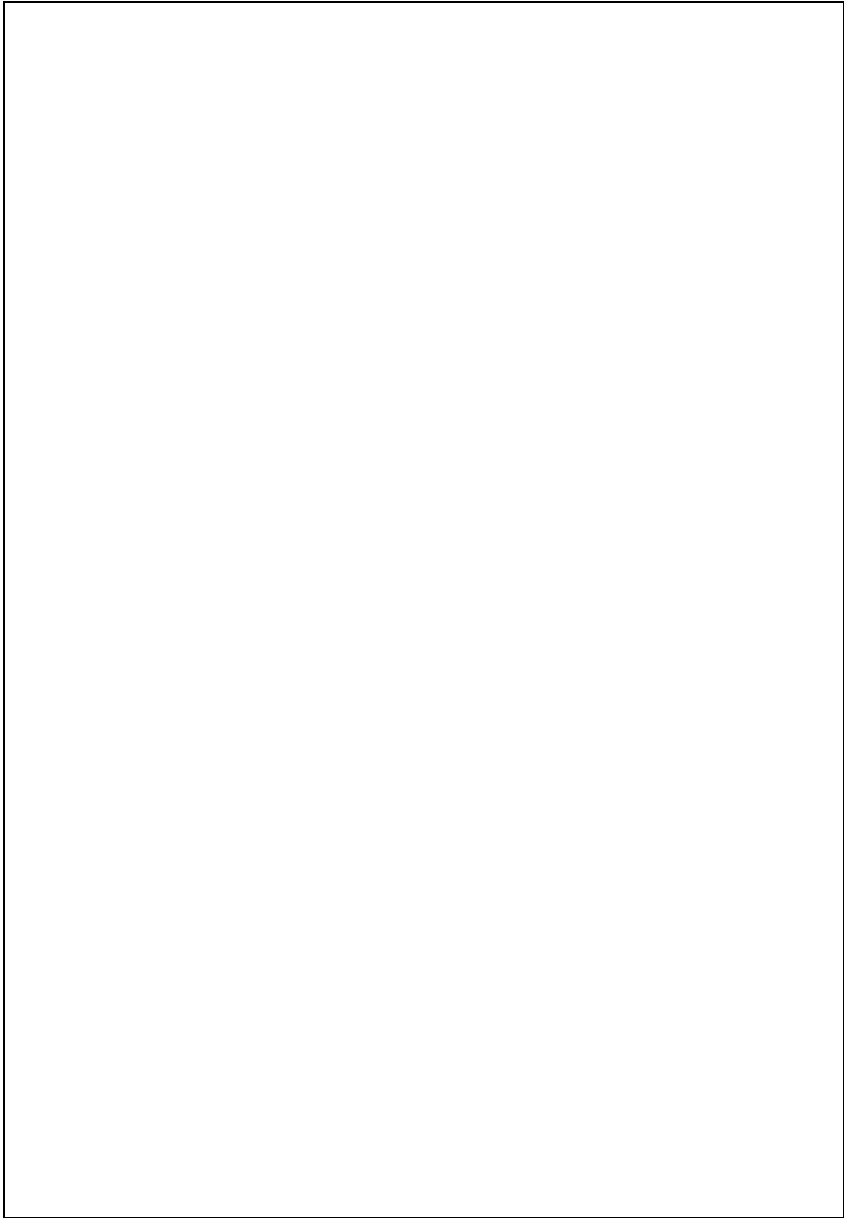


Figure 3. Concentrations of nitrate, ammonia and phosphate in groundwater near the Blue Heron active tile bed. At the time of sampling, the bed had been in operation for 18 years.

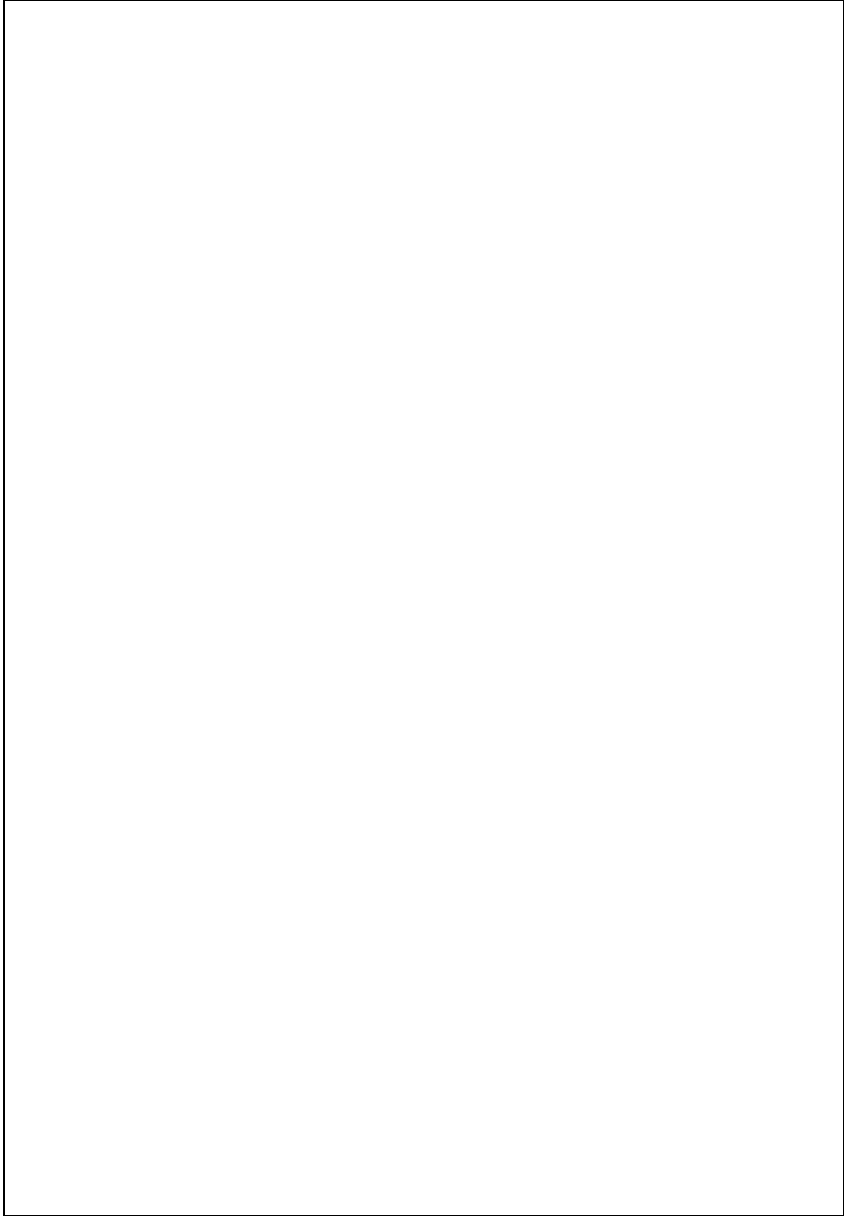


Figure 4. Concentrations of nutrients in the groundwater near the Camp Henry tile bed. The bed was active for 16 years at the time of sampling.

## Formerly-Developed Area

At the site where records show the presence of vault toilets, numerous buildings, and possibly a tile bed, elevated concentrations of  $\text{PO}_4$  and  $\text{NH}_3$  are present in the groundwater zone (Figure 5,  $\text{NH}_3$  not shown; Figure 6). Very high concentrations of  $\text{PO}_4$  (1-2 mg  $\text{PO}_4$ -P) were observed in isolated pockets close to locations of earlier discharges (Thompson et al., 1997). There are also very high concentrations of  $\text{PO}_4$  in isolated pockets along the edge of the marsh (> 8 mg/L  $\text{PO}_4$ -P), close to the former location of a vault toilet.

The natural anaerobic degradation of organic matter results in the release of  $\text{PO}_4$  and  $\text{NH}_3$  into groundwater. The elevated concentrations of nutrients observed close to the marsh may be a result of this natural process or a result of past wastewater release to the subsurface. The highest concentrations were observed close to wastewater release sites. Elsewhere along the groundwater/marsh interface concentrations are much lower than in this zone where previous releases were known to have occurred. These results suggest that the elevated concentrations observed along the marsh edge away from the sewage disposal areas are a result of the degradation of natural organic matter, and the higher concentrations are the result of residual sewage. Therefore, it appears that even after 20 years, there is the potential for sewage-derived  $\text{PO}_4$  and  $\text{NH}_3$  to persist in the groundwater zone.

## Discharge into Marsh

Hydraulic head and seepage meter measurements indicate groundwater flow is directed into the marsh for most of the year. For example, in May 1996 groundwater was observed to be discharging into the marsh at a rate of 0.02 and 0.13 litres/ $\text{m}^2$ /day (Ptacek et al., 1997). Even higher discharge rates are expected to occur during periods of high infiltration (spring melt, autumn rains) or in response to large declines in the elevation of the marsh.

## Implications

The results from the sampling at the active tile beds indicate that elevated concentrations of  $\text{NO}_3$ ,  $\text{NH}_3$ , and  $\text{PO}_4$  are present in the groundwater zone as a result of recent sewage discharge to the subsurface. Sampling at the site two years following decommissioning indicates concentrations of  $\text{NO}_3$  and  $\text{NH}_3$  declined, but high concentrations of  $\text{PO}_4$  remain. At a location where sewage release ceased two decades ago, concentrations of  $\text{PO}_4$  and  $\text{NH}_3$  are elevated, suggesting there is a long-term potential for nutrient persistence.

Geochemical monitoring at the marsh interface shows elevated concentrations of nutrients in the groundwater zone near the marsh edge and in the marsh pond bottom sediments. Hydraulic head measurements show groundwater flows toward the marsh for most of the year. Groundwater seepage measurements indicate seepage of groundwater into the marsh for most of the year.

Many Parks in Ontario rely on subsurface disposal of wastewater. Groundwater plumes emanating from these sites typically contain elevated concentrations of nutrients. These plumes can discharge into surface water bodies increasing the nutrient pool. After tile bed abandonment, release of phosphate from previously accumulated solid-phases can lead to additional release to surface water bodies.



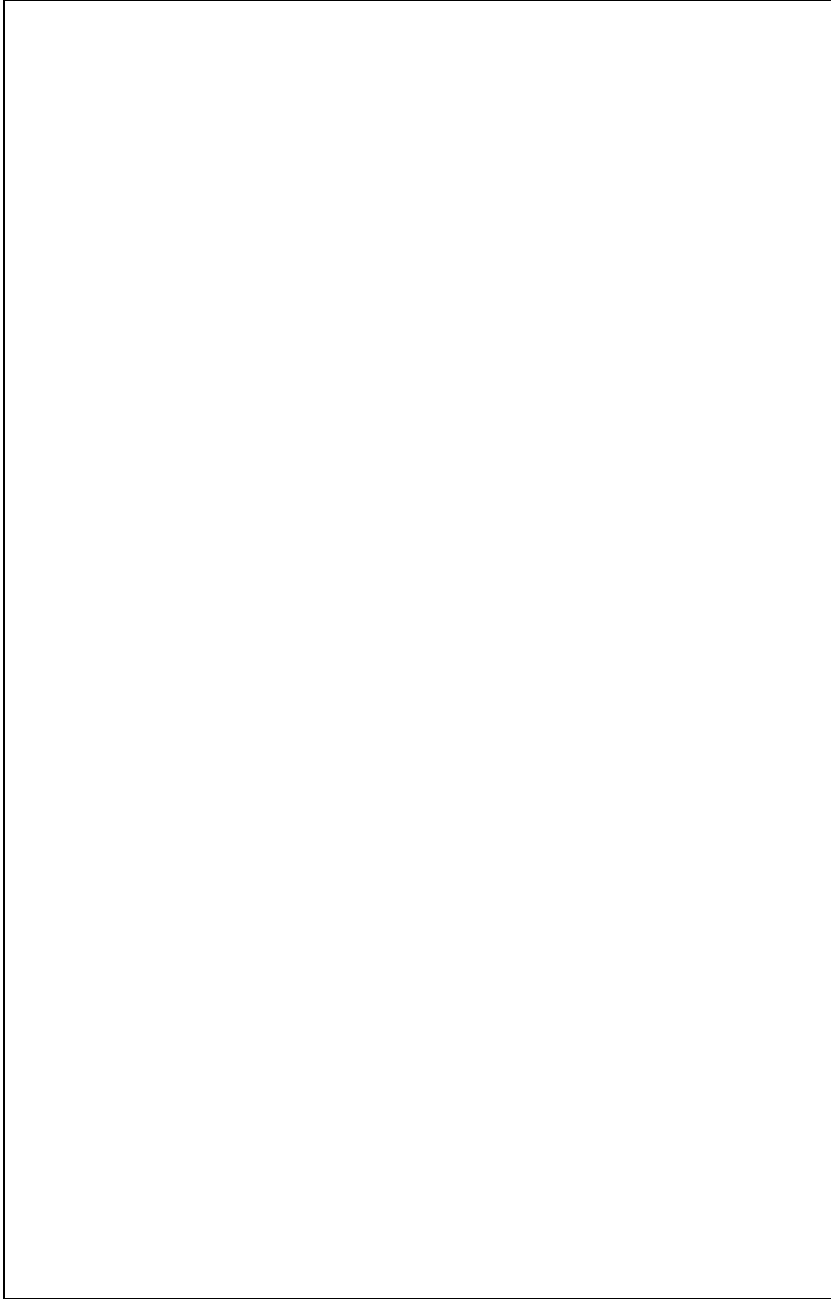


Figure 5. Concentrations of nutrients in the groundwater two years after the Camp Henry tile bed was decommissioned.

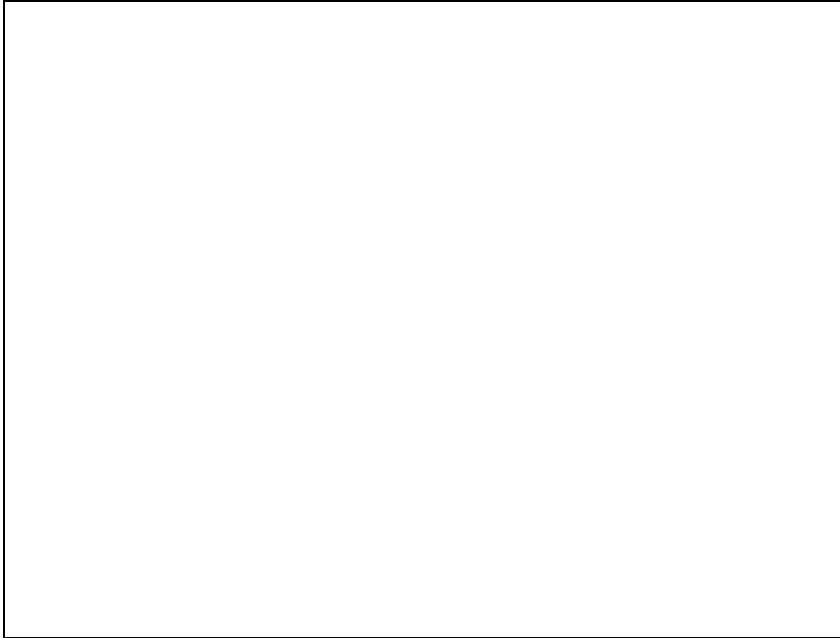


Figure 6. Concentrations of phosphate in area where on-site wastewater disposal last occurred 20 years ago (central area). Also shown are the locations of two active tile beds (top and bottom). This area had numerous buildings, a barn, vault toilets and possibly a tile bed prior that were removed by the Park

Sewage effluent typically contains between 5 and 15 mg/L P and loadings of 10,000 L/day are common at large park comfort stations. These comfort stations operate for many decades, therefore the total phosphorous released can be on the order of 100's to 1000's of kg over the life of a bed. Typical groundwater concentrations of  $\text{PO}_4$  are in the range of 0.1 to 3.0 mg/L P. Even if only a portion of this phosphorous is eventually leached at these concentrations, the loadings can be sufficiently large to cause a significant increase in the nutrient pool of marsh ponds. Once nutrients are present in the marsh water column and sediments, a portion will be buried, but a portion can also be regenerated each year for renewed biomass growth.

### Acknowledgements

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